

## FERTILIZATION INCREASES SOIL ORGANIC CARBON STOCKS BUT DOES NOT MITIGATE CLIMATE CHANGE IN THE ARGENTINE PAMPAS

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### ABSTRACT

Soils can act as carbon sinks mitigating global warming. This generates interest in determining how agricultural practices affect the stock of soil organic carbon (SOC). Fertilization increases SOC stocks and its impacts have been calculated at a global level. The aim of this study was to determine if fertilization leads to carbon sequestration in the Pampas of Argentina. A meta-analysis of local studies was performed to determine how fertilization impacts SOC stocks in agricultural soils. Forty studies were compiled presenting data from 27 experiments from which 109 comparisons of SOC stocks between fertilized and unfertilized treatments were extracted. Fertilization caused a significant increase of ca. 3% in the topsoil (0-20 cm) SOC stock. When nutrients application rate could be accurately defined ( $n=71$ ), the average increase in SOC stock was ca.  $1 \text{ t ha}^{-1}$ . Increments were much smaller in comparison with those reported in literature and the carbon footprint of applied nutrients ( $2.31 \text{ t Ceq. ha}^{-1}$ ). Even if increments in SOC stock in the subsoil were similar to those measured in the topsoil, greenhouse gas emission from fertilizer application would not be offset by the carbon sequestered in the soil. Fertilization has a low impact on SOC stocks and does not compensate greenhouse gases emissions in the Pampas. Rates of SOC increments resulting from fertilization identified in other regions of the world should not be applied in this region.

**Keywords:** carbon sequestration, life cycle assessment, management practices.

## LA FERTILIZACIÓN AUMENTA EL CARBONO ORGÁNICO DEL SUELO EN LA PAMPA ARGENTINA PERO NO MITIGA EL CAMBIO CLIMÁTICO

### RESUMEN

Los suelos pueden actuar como sumideros de carbono y mitigar el calentamiento global. Esto genera interés en determinar cómo las prácticas agrícolas afectan las reservas de carbono orgánico del suelo (COS). La fertilización produce aumentos en el COS y se han calculado coeficientes de su impacto a nivel global. El objetivo fue determinar si la práctica de la fertilización conduce al secuestro de carbono en la Región Pampeña de Argentina. Se realizó un metanálisis de estudios locales para determinar cómo la fertilización afecta el COS en los suelos agrícolas. Se compilieron cuarenta estudios que presentaban datos de 27 experimentos de los cuales se extrajeron 109 comparaciones de existencias de COS entre tratamientos fertilizados y no fertilizados. La fertilización determinó un aumento significativo de ca. 3% en la reserva de COS de la capa superior del suelo (0-20 cm). En los casos en los que la dosis de nutrientes aplicados estuvo bien definida ( $n=71$ ), el aumento promedio en el COS fue de ca.  $1 \text{ t ha}^{-1}$ . Este aumento fue mucho menor que los aumentos de COS informados en la literatura y que la huella de carbono de los nutrientes aplicados ( $2,31 \text{ t Ceq. ha}^{-1}$ ). Incluso suponiendo que el aumento relativo del COS en el subsuelo

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Recibido:  
04-09-23

Recibido con revisiones:  
25-10-23

Aceptado:  
27-10-23

fuera similar al medido en la capa superficial, las emisiones de gases de efecto invernadero procedentes de la aplicación de fertilizantes no serían compensadas por el carbono secuestrado en el suelo. La fertilización es una práctica que tiene bajo impacto en el COS y no mitiga el calentamiento global en la Región Pampeana. En esta región no se deben aplicar los coeficientes de aumento de carbono por fertilización determinados en otras regiones del mundo.

**Palabras clave:** secuestro de carbono, análisis de ciclo de vida, prácticas de manejo.

## INTRODUCTION

It is possible to transform cultivated soils into atmospheric carbon sinks as a mechanism to mitigate global warming (Paustian et al., 2016; Lal, 2017). With adequate management practices, around 20-35% of human carbon emissions could be sequestered in these soils (Minasny et al., 2017). Fertilization is among the management practices that allow increasing soil organic carbon (SOC) stock. Several global meta-analyses have shown that the application of nutrients to the soil promotes an increase in SOC stocks compared to unfertilized controls (Geisseler & Scow, 2014; Han et al., 2016; Ladha et al., 2011). The increase of the SOC stocks is produced by higher residue carbon input (Lu et al., 2011; Tian et al., 2015) and a decrease in microbial respiration (Ramirez et al., 2012; Xu et al., 2020) which leads to a longer residence time of the SOC (Lu et al., 2011). The increase of SOC content is produced only in agroecosystems with residue returned (Alvarez, 2005). A regional survey in China's croplands showed that SOC stock increased between 1980 and 2011 (Tao et al., 2019), in line with existing meta-analyses. Such increase of the SOC stock was partially attributed to a greater productivity of fertilized crops.

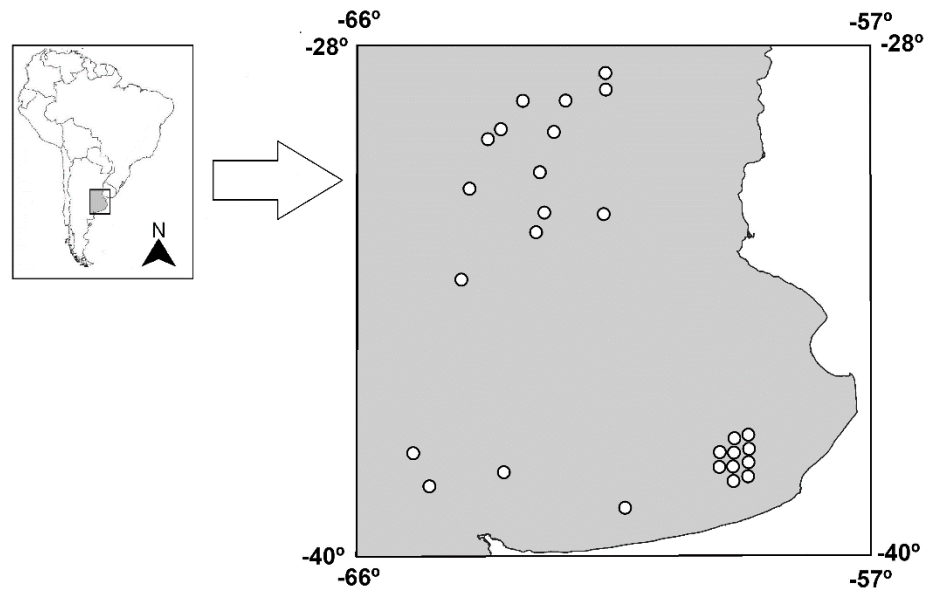
Unfortunately, not every increase in SOC stock implies that the greenhouse effect is mitigated. Life cycle studies show that the increases in SOC content due to improved management practices may not compensate for the effect of additional emission of greenhouse gases associated to the implementation of such practices (Schlesinger & Amundson, 2018). It is important that soils do not act as a source of CO<sub>2</sub> and become carbon sinks (Sykes et al., 2020), not only because the carbon cycle will be impacted but also soil productivity (Soussana et al., 2019) and quality (Fazera, 2012; Feller et al., 2012) will be improved. Consequently, to determine the effect of fertilization on global warming, its impact on SOC stock must be determined along with the associated greenhouse gas emissions.

The Pampas Region in Argentina is an area of great importance due to its capacity to produce grain crops (Dominguez & Rubio, 2019). Fertilization is a massively adopted practice but its effects on SOC stocks and greenhouse gas emissions are unknown. The objectives of this study are to evaluate the impact of fertilization on SOC levels in Pampean soils and to contrast the carbon sequestered in soils with the additional greenhouse emissions due to the implementation of such practice. This study seeks to assess whether the application of fertilizers can be considered a way to mitigate the greenhouse effect or a management practice that enhances global warming.

## MATERIALS AND METHODS

### Study area

The Pampas is a plain area of ca. 50 Mha, located between 28°S-40°S and 57°W-66 °W in Argentina (Figure 1). Natural vegetation consists of grassland and around 33 Mha are under cultivation (INDEC, 2022). The mean annual temperature varies from 14°C to 23°C from South to North and mean annual rainfall increases from 500 mm to 1200 mm from West to East. Mollisols are the predominant soils with both texture and SOC content following the rainfall gradient: coarse soils with low SOC content in the West to fine soils with high SOC content in the East (Berhongaray et al., 2013). Well-drained soils are cultivated with grain crops while hydromorphic areas have remanent grasslands used for grazing. Main crops are soybean (*Glycine max* L.) Merr.), wheat (*Triticum aestivum* L.), corn (*Zea mays* L.) and sunflower (*Helianthus annuus* L.) (MINAGRI, 2022). Alfalfa (*Medicago sativa* L.) based pastures are rotated with grain crops in some areas. Approximately 90% of the cultivated surface is managed with no-tillage (Noceli Pac, 2018) and crop residues are left in the field.



**Figure 1:** Map of the Pampas Region showing the location of experimental sites. Some points have been slightly shifted from their actual location to avoid overlapping.

**Figura 1:** Mapa de la Región Pampeana mostrando la ubicación de los sitios experimentales. Algunos puntos han sido levemente corridos de su ubicación real para evitar superposiciones.

#### Data search and processing

A bibliographic search was carried out between June 5 and June 10 of 2022 to identify articles where the impact of fertilizer application on SOC was evaluated in the Pampas region. The search period was set from 1980 to June 2022. Both the Google Scholar and Scopus databases were used to locate articles published in international scientific journals. The search terms were “Argentina” AND “fertilization” or “Argentina” AND “soil carbon”. In addition, an online local journal on soil science (Ciencia del Suelo) was fully reviewed. Proceedings of the Argentina’s National Soil Science Congress in which full length papers were published were reviewed (11 proceedings since year 2000) as well. An online search of technical bulletins of INTA, a governmental institution dedicated to technological improvement in the agricultural sector, was also performed. As a result of this search, 40 articles were identified in which SOC levels were investigated under different fertilization regimes. The published data corresponded to 27 field experiments distributed in the Pampas under a very wide range of soil and climate conditions (Table 1, Figure 1). To be included in this study an experiment had to meet the following conditions: 1) it should be a field experiment, 2) it should include a non-fertilized control compared with one or more fertilized treatments, 3) the SOC stock or concentration should be reported for all treatments, 4) the sampling depth should be specified, 5) the nutrient combinations of fertilizers should be defined, 6) the duration of the experiments should be indicated and 7) the experimental design should be clear and the number of replications mentioned.

Table 1: Main characteristics of the experiments used in the meta-analysis.

Tabla 1: Principales características de los experimentos incluidos en el meta-análisis.

Reference	Experiment	Location	Soil Type	Sampling depth (cm)	Years	Replications	Rotation	Sand (%)	Rainfall (mm)	Treatments	N rate (kg ha <sup>-1</sup> yr <sup>-1</sup> )	P rate (kg ha <sup>-1</sup> yr <sup>-1</sup> )	S rate (kg ha <sup>-1</sup> yr <sup>-1</sup> )
Ciampiti et al. 2011 García et al. 2010	1	Canals	Entic Haplustoll	0-20	6	3	C-W/S	35.4	920	NPS	Undefined	34	Undefined
Vega Jara et al. 2020 García et al. 2010 Ciampiti et al. 2011 Verdenelli et al. 2018	2	Teodelina	Typic Hapludoll	0-20	13	3	C-W/S	35.1	1110	PS NS NP NPS NPS+M	0 Undefined Undefined Undefined Undefined	37 0 37 37 37	21 21 0 21 21
Vega Jara et al. 2020 García et al. 2010 Ciampiti et al. 2011	3	Santa Isabel	Typic Argiudoll	0-20	13	2	C-W/S	20	967	PS NS NP NPS NPS+M	0 Undefined Undefined Undefined Undefined	37 0 37 37 37	21 21 0 21 21
Vega Jara et al. 2020 García et al. 2010 Ciampiti et al. 2011	4	Gral. Baldissera	Typic Hapludoll	0-20	13	3	C-S-W/S	28.1	920	PS NS NP NPS NPS+M	0 Undefined Undefined Undefined Undefined	37 0 37 37 37	21 21 0 21 21
Vega Jara et al. 2020	5	Amstrong	Aquic Argiudoll	0-20	13	3	C-S-W/S	31	1047	PS	0	37	21

García et al. 2010										NS	Undefined	0	21
Ciampiti et al. 2011										NP	Undefined	37	0
										NPS	Undefined	37	21
										NPS+M	Undefined	37	21
Vega Jara et al. 2020	6	San Jorge	Typic Argiudoll	0-20	13	3	C-S-W/S	30	1047	PS	0	37	21
García et al. 2010										NS	Undefined	0	21
Ciampiti et al. 2011										NP	Undefined	37	0
										NPS	Undefined	37	21
										NPS+M	Undefined	37	21
Martinez et al. 2019	7	Arequito	Typic Argiudoll	0-20	4	3	S-S	4.4	1003	PS	0	24	26
Martinez et al. 2014a							S-CC			NPS	34	24	26
Martinez et al. 2014b													
Crespo et al 2021a	8	Balcarce	Typic Argiudoll	0-20	10	3	S-S	42.9	887	PS	Undefined	Undefined	Undefined
Crespo et al 2021b							S-CC			NPS	Undefined	Undefined	Undefined
Beltrán et al. 2018													
Martinez et al. 2019													
Martinez et al. 2014a													
Martinez et al. 2014b													
Crespo et al 2021a	9	Marcoz Juarez	Typic Argiudoll	0-20	10	3	S-S	13.3	925	PS	Undefined	Undefined	Undefined
Crespo et al 2021b							S-CC			NPS	Undefined	Undefined	Undefined
Crespo et al 2021a	10	Paraná	Aquic Argiudoll	0-20	10	3	S-S	17.8	1060	PS	Undefined	Undefined	Undefined
Crespo et al 2021b							S-CC			NPS	Undefined	Undefined	Undefined

Crespo et al 2021a	11	General Villegas	Typic Hapludoll	0-20	10	3	S-S	65	909	PS	Undefined	Undefined	Undefined
Crespo et al 2021b							S-CC			NPS	Undefined	Undefined	Undefined
Melchiori et al. 2014	12	Paraná	Aquic Argiudoll	0-10	18	3	C-C	4.5	1060	N	69	0	0
										N	138	0	0
										N	276	0	0
Manso, Forjan, 2014	13	Barrow	Petrocalcic Paleudoll	0-15	10	6	C-S-W- Su-W-C- S-B/S- W-Su	7.7	911	N	Undefined	0	0
Vivas et al. 2012	14	Bernardo de Irigoyen	Typic Argiudoll	0-20	11	4	W/S-C-S	1.9	1047	P	0	20	0
										P	0	40	0
										S	0	0	12
										S	0	0	24
										S	0	0	36
Irizar et al. 2006	15	Pergamino	Typic Argiudoll	2500 <sup>1</sup>	25	2	W/S-C	12.5	1034	N	90	0	0
Cazorla et al. 2017	16	Marcos Juarez	Typic Argiudoll	0-18	13	3	C-W/S	6	925	N	88	0	0
Arrigo et al. 1993										N	176	0	0
Landriscini et al. 2020	17	Marcos Juarez	Typic Argiudoll	0-20	9	3	S-C	6	925	Undefined	Undefined	Undefined	Undefined
							S-C-CC <sub>1</sub>			Undefined	Undefined	Undefined	Undefined
							S-C-CC <sub>2</sub>			Undefined	Undefined	Undefined	Undefined
							S-C-CC <sub>3</sub>			Undefined	Undefined	Undefined	Undefined

							S-C-CC <sub>4</sub>			Undefined	Undefined	Undefined	Undefined
							S-C-CC <sub>5</sub>			Undefined	Undefined	Undefined	Undefined
Minoldo et al. 2008	18	Bordenave	Entic Haplustoll	0-20	31	4	W-W	77.3	614	NP	29	5	0
							W-CC			NP	29	5	0
							W-W-C-			NP	29	5	0
							C-CC			NP	29	5	0
							W-P			NP	29	5	0
Duval et al. 2019	19	Bordenave	Entic Haplustoll	0-20	13	3	W-W	77.3	614	NP	64	16	0
Miglierina et al. 2000						3	W-W			NP	64	16	0
Galantini et al. 2006						3	W-G			NP	32	8	0
Galantini et al. 2014													
Duval et al. 2010													
Landriscini et al. 2016													
Fabrizi et al. 1998	20	Balcarce	Typic Argiudoll	0-18	7	3	W-W	41.1	887	N	120	0	0
							W-W			P	0	22	0
							W-W			NP	120	22	0
							W-Su			N	60	0	0
							W-Su			P	0	11	0
							W-Su			NP	60	11	0
Studdert, Echeverría 2000	21	Balcarce	Typic Argiudoll Petrocalcic Paleudoll	0-17	12	4	W-W-W		887	N	105	0	0
Studdert et al. 2011							W-S-W	41.1		N	63	0	0
							W-Su-W			N	63	0	0

							W-C-W		N	93	0	0	
							S-W-W		N	73	0	0	
							S-S-W		N	38	0	0	
							S-Su-W		N	38	0	0	
							S-C-W		N	38	0	0	
							Su-W-W		N	73	0	0	
							Su-S-T		N	73	0	0	
							Su-Su-		N	73	0	0	
							-W		N	73	0	0	
							Su-C-W		N	55	0	0	
							C-W-W		N	93	0	0	
							C-S-W		N	58	0	0	
							C-Su-W		N	58	0	0	
							C-C-W		N	60	0	0	
			Typic										
			Argiudoll/										
Nontiel et al. 2019	22	Balcarce		0-20	14	3	W-P <sub>1</sub>	887	N	26	0	0	
			Petrocalcic										
			Paleudoll										
Studdert 2008					26	3	W-P <sub>2</sub>	41.1	N	49	0	0	
Eiza et al. 2004					26	3	W-P <sub>3</sub>		N	68	0	0	
González et al. 2012					26	3	W-P <sub>4</sub>		N	83	0	0	
					26	3	W-P <sub>5</sub>		N	83	0	0	
					14	3	W-P <sub>6</sub>		N	Undefined	0	0	
					14	3	W-P <sub>7</sub>		N	Undefined	0	0	
							W-S-W-						
Fabrizzi et al. 2003	23	Balcarce	Petrocalcic	0-15	7	4	-C-C-	41.1	887	N	120	0	0
			Paleudoll				-Su-C			N	120	0	0



Vidaurreta et al. 2012	24	Balcarce	Typic	0-20	10	3	C-S-W/	41.1	887	P	0	30	0	
			Argiudoll				S-W/S-							
			Petrocalcic				C-S-							
			Paleudoll				W/S-C-							
							S-W/S			P	0	30	0	
Studdert et al. 2015	25	Balcarce	Typic	0-20	22	6	C-Su-W	41.1	887	N	127	0	0	
			Argiudoll											
			Petrocalcic											
			Paleudoll											
Dominguez et al. 2009														
Moreno et al. 2014														
Diovisalvi et al. 2008														
Diovisalvi et al. 2006														
Studdert et al. 2006														
Divito et al. 2011	26	Balcarce	Typic	0-20	8	8	C-S-W/S	41.1	887	N	50	0	0	
			Argiudoll											
			Petrocalcic											
			Paleudoll											
										N	91	0	0	
Wyngaard et al. 2012	27	Balcarce	Typic	0-20	8	3	C-S-W/S	41.1	887	NP	86	38	0	
			Argiudoll											
Wyngaard et al. 2013										NS	86	0	19	
										PS	0	38	19	
											NPS	86	38	19
											NPS+M	86	38	19
											NPS+M+L	86	38	19
											NP	86	38	0

NS	86	0	19
PS	0	38	19
NPS	86	38	19
NPS+M	86	38	19
NPS+M+L	86	38	19

C= corn, W= wheat, S= soybean, W/S= double crop in a year wheat and soybean, Su= sunflower, CC= cover crop, B= barley, B/S= double crop barley and soybean, P= pasture. G= grassland.

N= nitrogen, P= phosphorus, S= sulfur, M= micronutrients, L= lime

1= soil equivalent mass (t ha<sup>-1</sup>)

Information regarding soil type and texture and total rainfall at the experimental sites was reported in the articles or was taken from soil maps (GeoINTA, 2022) and from the Climate Research Unit (CRU) database (CRU, 2022). To avoid double counting of data when several published articles involved the same experiment, only data from the longest period were used. Data were mostly taken from tables but data acquisition software was used to access graphical information (Getdata Graph Digitizer 2.24) when necessary. Since the study focused on the effects of fertilization on SOC sequestration, SOC stocks rather than concentrations were compared under fertilized and unfertilized conditions. If papers reported SOC stock ( $\text{t ha}^{-1}$ ), such data were directly used for analysis. If carbon concentration and soil bulk density were available, SOC stock was calculated. When only SOC concentration was reported, bulk density was estimated as described by Post and Kwon (2000) from SOC concentration and using  $1.64 \text{ g cm}^{-3}$  as bulk density of the mineral soil fraction. A factor of 1.72 was used for converting organic matter into carbon (Nelson & Sommers, 1996) in some cases.

Average SOC stock for each treatment was extracted and 109 fertilized vs. unfertilized pairs of SOC stocks were calculated for comparisons (Tables 2, 3). Sampling depth was 0-20 cm in about 80% of the experiments. In the remaining trials, soil was sampled at a lower depth, but those data were still included in the meta-analysis. All data were considered to come from topsoil. The initial SOC level of the soil was also registered when available. In all cases, the number of replications was available but standard deviations could be extracted in only 25% of the cases. In these latter cases, the average standard deviation/mean ratio was 0.068. An imputation procedure was used to estimate standard deviation of the whole dataset assuming that standard deviations were equal to 6.8% of the means (Wiebe et al., 2006). In some experiments, the use of the same unfertilized control treatment against which several fertilized treatments were compared to generates non-independent data in a meta-analysis, thus increasing the probability of Type I Error (Lajeunesse, 2011). In experiments from which various data pairs were extracted, SOC stocks of the fertilized treatments were averaged to achieve independent data (Noble et al. 2017, Song et al. 2020). The number of replicates of the fertilization treatments that were averaged was the sum of the replicates of the individual treatments. In such scenario, 69 independent data pairs were used for analysis. Although in some experiments, nutrients rates were not accurately defined, 71 experiments included nitrogen, phosphorus and sulfur rates, so cumulative rates applied to the soil could be calculated. When graminaceous crops were included, the ratio between the number of wheat, corn or graminaceous cover crops over the total number of crops included in the rotation was calculated.

**Table 2:** Main data extracted from 27 experiments with 109 fertilization treatments in the Pampas regions. Unweighted averages and ranges are shown.

**Tabla 2:** Principales datos extractados de 27 experimentos con 109 tratamientos de fertilización. Se muestran promedios no ponderados y rangos.

Variable	N	Unit	Mean	Minimum	Maximum
Duration	33 <sup>1</sup>	(years)	14.1	4	31
Depth	27	(cm)	18.8	10	20
Sand	27	(%)	30.9	1.9	77.3
Rainfall	27	(mm)	923	614	1110
Graminaceous crops	68 <sup>2</sup>	(%)	59	0	100
Initial carbon	25 <sup>3</sup>	(t ha <sup>-1</sup> )	62.6	30.5	92.7
Carbon in control	68 <sup>4</sup>	(t ha <sup>-1</sup> )	53.8	23.7	77.8
Carbon loss in control (SOC Initial – SOC control)	47	(t ha <sup>-1</sup> )	6.20	-12.2 <sup>5</sup>	17.7
Annual carbon loss in control	47	(t ha <sup>-1</sup> yr <sup>-1</sup> )	0.45	-2.03 <sup>5</sup>	1.44
Carbon in fertilized treatment	109	(t ha <sup>-1</sup> )	54.1	22.8	79.9
Carbon loss in fertilized treatment (SOC Initial – SOC fertilized)	81	(t ha <sup>-1</sup> )	4.00	-16.5 <sup>5</sup>	27.4
Annual carbon loss in fertilized treatment	81	(t ha <sup>-1</sup> yr <sup>-1</sup> )	0.38	-2.23 <sup>5</sup>	3.43
Carbon gain in fertilized treatment (SOC fertilized – SOC control)	109	(t ha <sup>-1</sup> )	1.47	-15.4	11.6
Annual carbon gain in fertilized treatment	109	(t ha <sup>-1</sup> yr <sup>-1</sup> )	0.12	-1.93	0.89

<sup>1</sup> In one experiment, fertilization treatments started at different times.

<sup>2</sup> Sixty eight different rotations were tested in 27 experiments.

<sup>3</sup> In some experiments initial carbon content data were not available.

<sup>4</sup> In some experiments various control treatments were used.

<sup>5</sup> Negative data indicate carbon gain.

**Table 3:** Summary of fertilization treatments tested in 27 field experiments and the corresponding unweighted mean difference (SOC fertilized – SOC control).

**Tabla 3:** Resumen de los tratamientos de fertilización testeados en 27 experimentos y sus correspondientes diferencias medias no ponderadas (COS fertilizado – COS control).

Treatments	n	Mean N-P-S rates (kg ha <sup>-1</sup> )	Unweighted mean difference (t C ha <sup>-1</sup> )
All (defined and undefined rates)	109		1.47
Defined rates	71		0.97
Only N	34	85-0-0	1.73
Only P	6	0-26-0	-0.64
Only S	3	0-0-24	1.90
NP	11	57-15-0	1.20
NS	2	86-19-0	2.90
PS	8	0-36-21	0.63
NPS	7	79-36-20	2.30

**Meta-analysis**

Meta-analytic methods were used to analyze the data. A detailed description of models and equations used can be found elsewhere (Alvarez, 2021). Two different weighting functions were applied: the inverse of the pooled variance (Rosenberg et al. 2000) and the sample size (Adams et al., 1997). The response ratio (*RR*) was chosen as effect size (ratio of SOC in fertilized treatment/SOC unfertilized control) which was log-transformed to approach the normal distribution before statistical analysis (Hedges et al., 1999). A bias-corrected for skewness 95% confidence interval was estimated by bootstrapping methods (Adams et al., 1997). Results were presented as percent change ( $(RR-1)*100$ ). When the confidence interval did not overlap with 0 (zero), significant effects ( $P<0.05$ ) of the fertilizer treatment on SOC stock in relation to the control were recognized (Rosenberg et al., 2000). Average *RR* and confidence intervals were calculated using a random effect model (Rosenberg et al. 2000) assuming that effect sizes were not fixed across all studies (Gurevitch & Hedges, 1999). The software used was MetaWin 2.0 (Rosenberg et al., 2000). For comparative purposes, the mean difference (SOC fertilized – SOC control) was also calculated. Pearson's correlation analysis was used to study the effects of environmental variables on *RR*. No heterogeneity analysis to differentiate the effects of individual nutrients or rates on SOC could be conducted, since sample size in each category was insufficient (Kallenbach & Grandy 2011).

**Carbon footprint of fertilizers**

Carbon footprint due to the manufacture, transport and use of fertilizers has not been determined in Argentina. For this reason, literature values were used. Carbon footprint depends on the energy efficiency of the production process, the energy source used and the distance that the fertilizer is transported (Zhang et al., 2013). Average values were calculated as estimates for the Pampas, considering existing differences among regions (Table 4). When the difference between increments in SOC stock due to fertilization and the cumulative carbon footprint of fertilizers was negative, it was considered as a net flow of greenhouse gases to the atmosphere.

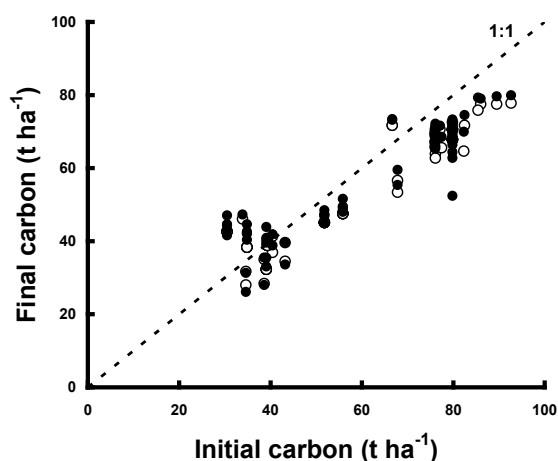
**Table 4:** Carbon footprint coefficients of fertilizer nutrients. In the cited literature carbon footprint is expressed as CO<sub>2</sub> emission and its carbon equivalence (Ceq.) was calculated.

**Tabla 4:** Huella de carbono de los fertilizantes. Se calculó la equivalencia en carbono (Ceq.) de la huella de carbono presentada en la literatura como CO<sub>2</sub>.

Nutrient	Reference	Region	Carbon footprints coefficient	
			CO <sub>2</sub> eq. (t t <sup>-1</sup> nutrient)	Ceq. (t t <sup>-1</sup> nutrient)
Nitrogen	Chojnacka et al. (2019)	Europe	9.3	2.51
	Zhang et al. (2013)	Europe	9.7	2.62
	Zhang et al. (2013)	China	13.5	3.65
	Brentrup et al. (2016)	IPCC methodology	10.4	2.81
	Mean		10.7	2.89
Phosphorus	Ledgard & Falconer (2019)	New Zealand	1.90	0.51
	Wood & Cowie (2004)	Europe	1.23	0.33
	West & Marland, (2002)	USA	1.44	0.39
	Mean		1.52	0.41
Sulfur	European Commission 2009	Europe	0.063	0.017

## RESULTS

Soil organic carbon measurements were performed in the topsoil, usually for the 0-20 cm in most experiments (Table 1). The average duration of the experiments was 14.1 years. In such period, most soils from unfertilized control and fertilized treatments lost SOC stock (Figure 2). Nonetheless, in three experiments where SOC initial values were low, SOC stocks increased, both in unfertilized and fertilized treatments. The mean loss of SOC stock was 6.2 t ha<sup>-1</sup> and 4.0 t ha<sup>-1</sup> for unfertilized control and fertilized treatments, respectively (Table 2). The average annual loss of SOC stock across the experiments varied between 0.45 and 0.38 t ha<sup>-1</sup> yr<sup>-1</sup> depending on whether it was the control or fertilized treatment.

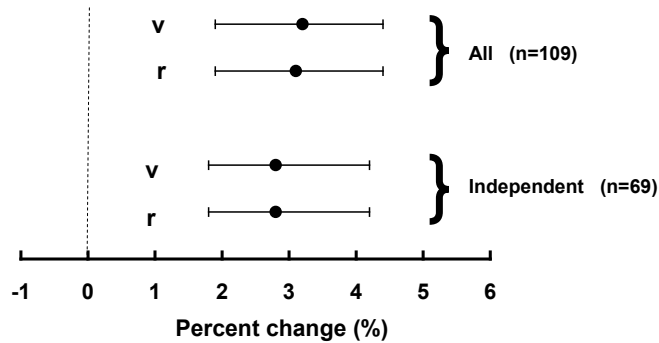


**Figure 2:** Relationship between initial and final SOC stocks of field experiments. Empty drops = unfertilized controls, full drops = fertilized treatments. Values below the 1:1 line indicate carbon loss. Values above the 1:1 line indicate carbon gain.

**Figura 2:** Relación entre el stock de COS al inicio y al final de los experimentos. Puntos vacíos= controles no fertilizados, puntos llenos= tratamientos fertilizados. Valores menores que la línea 1:1 indican pérdida de carbono y valores mayores indican ganancia de carbono.

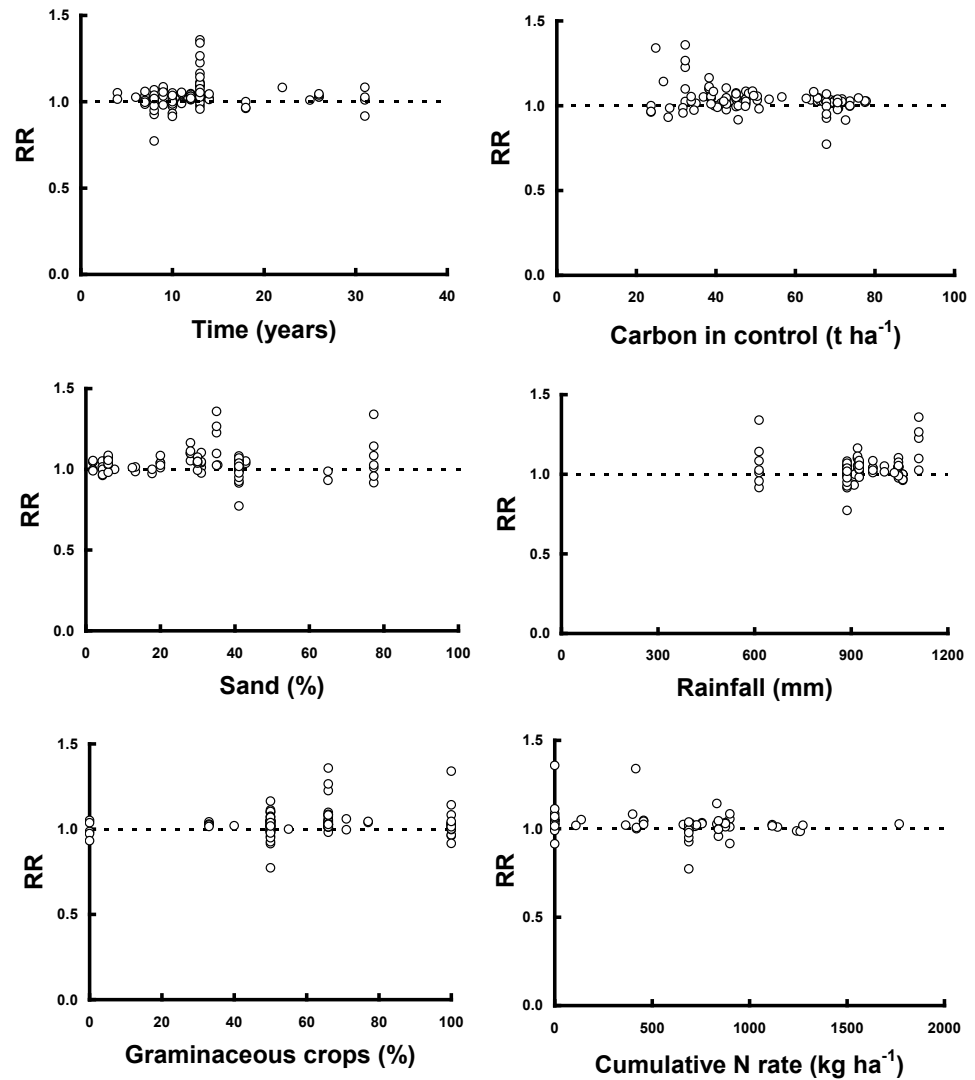
Most nutrient combinations produced increases in SOC stock compared to unfertilized control treatments

(Table 3). The meta-analysis indicated a significant SOC stock increase resulting from fertilization (Figure 3). An increase of about 3% of the SOC stock in the topsoil was reached, considering either the entire dataset or the independent dataset. A similar result was attained when weighting by variability or sample size. It was not possible to isolate controlling factors of the increase in SOC stock due to fertilization. The RR did not correlate with any of the following variables: duration of the experiment, soil texture, initial SOC content, rainfall at the experimental site, crop rotation or cumulative nitrogen rate (Figure 4).



*Figure 3:* Percent change of soil organic carbon stock in fertilized treatments as compared with unfertilized controls from the meta-analysis. v = weighted by variability; r = weighted by sample size. Analyses are shown for the entire dataset (All) and for the independent dataset (Independent).

*Figura 3:* Resultados del meta-análisis. Porcentaje de cambio del stock de COS en los tratamientos fertilizados relativos al control. v = ponderado por variabilidad; r = ponderado por tamaño de muestra. Se presentan dos análisis, uno para el set completo de datos (All) y otro solo para los datos independientes (Independent).



*Figure 4:* Correlation analysis between response ratio (RR) and management and environmental variables for 27 field experiments. No significant functions could be fitted.

*Figura 4:* Análisis de correlación entre la relación de respuesta (RR) y algunas variables ambientales y de manejo en 27 experimentos. No se pudo ajustar ninguna función significativa.

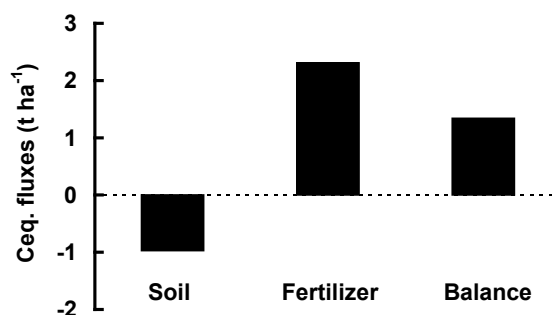
For 37% of the extracted fertilization treatments, the rates of the applied nutrients were not adequately defined. In some of these experiments the methods chosen to define the fertilizer rates were based on nutrient balances which usually leads to the application of high rates of nutrients, although they were not reported. The average rate of nitrogen was 3 to 4 times higher than the rates of phosphorus and sulfur (Table 5). The overall carbon footprint was much larger for nitrogen than for the other nutrients, due to both higher applied rates and larger greenhouse gas emissions of this nutrient (Table 5). Considering only those studies (71) in which the rates of all the applied nutrients could be calculated, the mean cumulative carbon footprint due to nitrogen, phosphorus and sulfur application was 2.31 t Ceq ha<sup>-1</sup>, nitrogen being responsible for approximately 98% of the emissions during the whole period of the experiments (Table 5).



**Table 5:** Cumulative carbon footprint of fertilization experiments in the Pampas.**Tabla 5:** Huella de carbono acumulada en los experimentos pampeanos.

Nutrient	n	Mean rate (kg nutrient ha <sup>-1</sup> yr <sup>-1</sup> )	Carbon footprints coefficient (t Ceq. t <sup>-1</sup> nutrient)	Cumulative rate (t nutrient ha <sup>-1</sup> )	Cumulative carbon
					footprint (Ceq. t ha <sup>-1</sup> )
Nitrogen	54	78.5	2.89	1.03	2.98
Phosphorus	32	26.8	0.41	0.27	0.11
Sulfur	20	21.0	0.017	0.19	0.0032

For the 71 treatments with defined nutrient rates it was possible to compare the cumulative SOC stock increase with the cumulative carbon footprint of fertilization. The fertilizer carbon footprint more than doubled SOC removal by fertilization (Figure 5). As a consequence, there was a net flux of greenhouse gases into the atmosphere by fertilization treatments.



**Figure 5:** Greenhouse gas fluxes related to fertilizer application to and from the atmosphere (calculated as carbon equivalent (Ceq.) units). Negative fluxes indicate SOC increases in fertilized treatments related to control. Positive fluxes are emission of Ceq. to the atmosphere due to fertilizer use. Balance is the difference between both previous fluxes.

**Figura 5:** Flujos de gases de efecto invernadero desde y hacia la atmósfera producidos por la fertilización (calculados como equivalentes de carbono, Ceq.). Los flujos negativos indican incrementos del COS en los tratamientos fertilizados en relación al control. Flujos positivos representan emisiones de Ceq. hacia la atmósfera debido al uso de fertilizantes. El balance es la diferencia entre ambos flujos.

## DISCUSSION

Based on published results from literature, this meta-analysis allowed determining that fertilization increases SOC stocks of the Pampas soils. Most global studies found considerable increases in SOC content in the upper soil layer due to fertilization, varying between 6% (Huang et al., 2020) and 15-16% (Geisseler & Scow, 2014; Han et al., 2016). Increments observed in the current study, however, were generally small (averaging 3%) and were not related to site or management factors, in agreement with findings from a prior global meta-analysis about the effects of nitrogen on cultivated soils (Lu et al., 2011).

The average carbon sequestration rate in fertilized soils in relation to unfertilized controls was 0.12 t ha<sup>-1</sup> yr<sup>-1</sup> with a total increase of 1.47 t C ha<sup>-1</sup> in the Pampas soils. This represents an annual increase of 2.2% of the SOC stock due to fertilization. The results of a large number of experiments in China show an average sequestration rate of ca. 0.15 t C ha<sup>-1</sup> yr<sup>-1</sup> (Waqas et al., 2020) with balanced fertilization. The total increase in SOC stock ranged from 1 t C ha<sup>-1</sup> in initially rich SOC soils to 6 t C ha<sup>-1</sup> in initially poor SOC soils (Ren et al., 2021). The effect of fertilization on SOC content in the Pampas is generally lower than that observed in most

synthesis studies mentioned above. The slight effect of fertilization on SOC stock in the Pampas Region seems to be related to the low rate of nutrients applied. When SOC stocks were calculated for different data sets, i.e. treatments that received either low ( $< 50 \text{ kg N ha}^{-1} \text{ year}^{-1}$ ;  $n = 11$ ) or high N rates ( $> 50 \text{ kg N ha}^{-1} \text{ year}^{-1}$ ;  $n = 30$ ), SOC increases of 1.01 and 1.99 t SOC  $\text{ha}^{-1}$  were identified, respectively. As suggested, SOC changes respond to the amount of nitrogen applied.

At a global scale, it has been estimated that nitrogen addition generates an increase in plant biomass (Xu et al., 2020) and residue input to soils ranging between 5 % (Xu et al., 2020) and 30-35 % (Lu et al., 2011; Huang et al., 2020). The enhanced residue input through fertilization rules the increase in SOC stocks, especially under reduced soil fertility (Huang et al., 2020). In China's croplands, yield increases of the main crops due to fertilization vary between 50% and 300% depending on the species (Waqas et al., 2020), which represents a huge difference in residue carbon inputs for fertilized or unfertilized soils. In the experiments across the Pampas Region included in this meta-analysis, the effect of fertilization on carbon input from crop residues was much less important. Carbon inputs from crop residues were evaluated in 13 of the experiments. Mean carbon inputs were only 12% higher in fertilized soils than in unfertilized controls (data not shown), which may account for the slight impact of fertilization on SOC stocks. Nevertheless, regular nutrient rates used by farmers of the Pampas are approximately half the rate applied in the experiments included in this meta-analysis (Alvarez et al., 2021). Therefore, the effect of fertilization on SOC stock under common agricultural management conditions is expected to be even lower than that estimated in this study.

Existing literature is contradictory as to what factors regulate the impact of fertilization on SOC stocks. It has been reported that temperature has positive (Ren et al., 2021), negative (Lu et al., 2011) or null (Han et al., 2016) effects on changes in SOC stocks produced by the application of nutrients. The impact of texture on SOC change is unclear (Waqas et al., 2020). Precipitation can have a positive impact (Ren et al., 2021) while the initial SOC level has a negative one (Ren et al., 2021). However, none of these possible effects have been observed in the Pampas experiments. The slight increase in SOC stocks of the Pampas soils cannot be attributed to a SOC level close to saturation. A regional analysis has shown that agricultural soils of the Pampas are far from saturation and they could double their SOC content in the topsoil (Alvarez & Berhongaray, 2021). Although fertilization led to increases in SOC stocks when comparing fertilized to unfertilized treatments, overall SOC stocks declined in most experiments compared to initial levels. The decrease of SOC stocks during the experiments was 4-6 t  $\text{ha}^{-1}$ , a larger amount than the gain produced by fertilization. Fertilization did not compensate for SOC losses due to cultivation in most cases.

Unfortunately, the impact of fertilization on the subsurface soil was not evaluated in any of compiled experiments. Since SOC has not been affected below 50 cm depth for over a century of cultivation in the Pampas (Berhongaray et al., 2013), it can be assumed that below that depth there will be no impact of fertilization on soil carbon pools. Under the most optimistic scenario, a similar relative increase in SOC can be assumed to occur in topsoil and subsoil. The average topsoil SOC stock for the 71 treatments in which nutrient rates were accurately defined was 57.5 t  $\text{ha}^{-1}$  and the average RR was 1.02. By applying a SOC stratification model (Berhongaray et al., 2013) for Pampas soils, the SOC stock for the 0-50 cm was estimated to be 96.1 t  $\text{ha}^{-1}$ . The potential carbon sequestration resulting from fertilization for the 0-50 cm soil layer would increase to 1.9 t  $\text{ha}^{-1}$ , but it would still be lower than the cumulative carbon footprint (2.31 t Ceq  $\text{ha}^{-1}$ ). Consequently, fertilization does not contribute to mitigation of climate change in the region.

Studies on carbon footprint in agriculture do not normally consider changes in soil carbon due to the difficulty of estimating them. For example, in the Pampas Region the carbon footprint of corn production has been calculated at 1.4 t  $\text{CO}_2 \text{ eq. ha}^{-1} \text{ yr}^{-1}$  (0.38 t Ceq.  $\text{ha}^{-1} \text{ yr}^{-1}$ ) in the Province of Córdoba, of which 40% results from the use of nitrogen fertilizers (Bongiovani et al. 2023). If the carbon sequestration coefficients in the soil by fertilization estimated in the present study were used, the carbon footprint of corn production would be reduced by 30%. This indicates the importance of having data on the effect of management practices such as fertilization on SOC stock. However, site-specific coefficients should be available. In the case of those calculated in this study, most of the experiments were carried out in humid areas and extrapolating the averages to semi-arid areas is dangerous.

## CONCLUSIONS

In the Pampas, small increases in SOC stocks are achieved through fertilization when comparing fertilized versus unfertilized soils. However, most soils in the experiments studied lost SOC over time and such loss was not compensated by the application of nutrients. At the same time, the greenhouse gases emitted due to the manufacture, transport and application of fertilizers will generate a warming effect that will not be compensated by carbon sequestration in the soil due to fertilization. Nutrient application should be considered as a practice with no potential to mitigate global warming in the region.

## ACKNOWLEDGEMENT

This study was granted by the University of Buenos Aires (UBACYT 20020170100016BA).

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